

Leslie Matrix Model For *Euphorbia Hirta* L Population

ASHA GUPTA

Department of Life Sciences
Centre of Advanced Study in Life Sciences
Manipur University, Canchipur-795003,
INDIA

Abstract: - Invasive species management has developed into a highly specialized field utilizing a systems approach. It requires knowledge of their life history, growth requirements, and population dynamics that integrate their biology and control. The foundation of strategic planning for the management of invasive species is laid by demographic studies, which record the birth, growth, reproduction, and death of individuals within a population. The present study makes use of the Discrete Leslie Matrix Model to analyze the growth in the age-structured population of *Euphorbia hirta*, an invasive species in agrosystems, identify critical stages in the species' life cycle, and project the structure and size of future population.

Key-Words: Invasive species, Leslie Matrix, critical stage, management, future population

Received: November 28, 2023. Revised: June 19, 2024. Accepted: July 12, 2024. Published: August 12, 2024.

1 Introduction

Invasive alien plant species are those that have been intentionally or unintentionally brought outside of their native habitat. They have an impact on human safety, habitats, biodiversity, ecology, and spread out of control [1; 2]. A biological invasion has been identified as one of the main causes of economic and environmental disruption, and biodiversity loss. Compared to native plants, invasive alien plants have advantages such as faster growth, greater photosynthetic rates, higher reproductive output, more biomass, lower carbon-to-nutrient ratios in tissue, stronger nutrient absorption capacities, and higher plasticity levels [3].

For efficient management and population regulation of invasive species, one must have a thorough understanding of the ecology, morphology, reproductive biology, physiology, and biochemistry of such species as a wide variety of factors regulate the density, growth, and competitive ability of these plants. According to Funk [4], invasive species have a greater strategic advantage for nutrient use over native plants. They are also more common along roadside and places with anthropogenic disturbances [5]. Plant invasions are detrimental to ecology and global biodiversity changing the landscapes. Many invasive weed species have invaded terrestrial crops worldwide [6] and drastically lowered agricultural output. Species invasions are a major component in global change resulting in habitat degradation, altering the biological diversity and environmental mechanism,

causing extinction of local flora and fauna, modifying ecosystem functioning and services, and promoting subsequent invasions that exacerbate the damage. [7;8;9; 10;11]. Climate change and biological invasions represent two of the largest threats to biodiversity in the Anthropocene [12]. Early detection of invasive plants, can help in weed management by efficient eradication and billions of dollars for ongoing control to stem biodiversity loss [13].

Adaptation of weed management strategies reduce the incidence of invasive species, decrease their undesirable effects, and optimize land use through the combination of preventive and control practices [14].

Invasive alien plant species may have a major effect on global agriculture, which continues to have an impact on food security worldwide [15]. Globally, invasive weed species need to be controlled via mechanical, physical, biological, and chemical methods. Many invasive plant species do not have any biological control agents [16]. Climate change [17], deforestation, ecological degradation, and anthropogenic disturbances all worsen agricultural production worldwide.

1.1 Demographic models

Mathematical models significantly contribute to the prediction the spread of invasive species and directing the optimal allocation of resources for their prevention, control, or eradication [18;19]. Models are indispensable tools for managing

invasive species. According to Baker and Bode [20], scientists can utilize them to calculate vital rates like the rate of spread, model the possible impacts of invasive species, and investigate the consequences of different strategies for controlling or eliminating them.

Introduced species may experience climatic niche shifts when moving to new continents because of changes to their fundamental or realized niche [21]. Whereas niche shifts are evaluated in 'climatic space'—usually with ordination techniques—the general goal of species distribution models (SDMs) is to project climatic niche models into geographic space using Matrix or other predictive models. Demographic models are useful for understanding population processes and stages in the life cycle of the species that could be most effectively targeted with management. The study carried out by Bogdan et al [22] in demographic data of an Israeli *Carpobrotus* population gives an Integral Projection Model and through the analysis of asymptotic growth rates and population sensitivities and elasticities demonstrated the population as stable, and reducing the survival of the largest individuals reduced the overall population growth.

Chung et al [23], developed an integrated spatial model to manage common ragweed (*Ambrosia artemisiifolia* var. *elatior* (L.) Decs) using various models, including species distribution BIOMOD2, landchange LCM, dispersal MigClim and optimization model prioritize and proposed a new 'removal effect index' for evaluation in time series.

Guetling et al [24], developed habitat susceptibility models (HSM) created within geographic information systems (GIS) to combine spatial environmental data at known infestations and predict areas likely to be invaded based on similar ground conditions [25] for meadow hawkweed and orange hawkweed. With known locations and environmental data, the predictive models were used to estimate habitat susceptibility for invaders by determining the indicator species.

Species distribution models (SDMs) are often used to produce risk maps to guide conservation management and decision-making with regard to invasive alien species (IAS). Davis et al [26] developed WiSDM, a semi-automated workflow to democratize the creation of open, reproducible, transparent, invasive alien species risk maps.

Worldwide, it is generally acknowledged that invasive species pose a serious threat to native biodiversity, ecosystem function, and economic interests at a global scale [27]. Structured population models (like matrix population models [28], integral projection models [29] of invasive plants provide a tool for producing comprehensive fitness estimates and identify sensitive vital rates (e.g. survival, growth) to target with management [28;30;31].

1.2 Population Dynamics of Invasive species

Invasive species management is concerned with maximizing mortality and lowering the reproduction and minimizing the loss of resources resulting from invasive species competition. In this context the role of natural and man-managed factors that regulate the size of invasive species population becomes of paramount importance. A classification of individuals by age in such a population provides reasonably accurate prediction of their demographic potential. Population dynamics of species helps in analyzing patterns related to growth, reproduction and mortality theoretically in mathematical terms. Demography studies help to learn empirically how the population grows in nature. They deal it by keeping the track of the birth, growth, reproduction and death of individuals in a population. It can then form the basis of strategic planning for invasive species control [32;33].

1.3 Matrix population models

Matrix population models are categorized into two types of models that are in vogue viz. Age structured models described by Leslie [34] and stage structured models described by Lefkovitch[35].

Sensitivity and elasticity assessments are also performed using matrix population models [36;37]. When an individual's attribute other than age is a stronger predictor of survival and reproduction, then according to Caswell [28] and Cochran and Ellner [38] the models that are preferred are stage structured matrix models.

In numerous research [39; 40;41; 42], vital rates in stage-structured population models can be inferred from age-structured vital rates, where age classes are grouped together to form a stage.

There has been an upsurge in the use of stage-structured population models, and a life table analysis is often used to estimate the vital rates in these models [43]. The relationship between the number of stages and different statistics derived from stage-structured population matrices is also covered by Salguero-Gomez & Plotkin [44]. Lebreton [45] suggests using models that are stage-

structured, meaning that Stages are embedded within the age classes. The dynamics of stage-structured populations have been described by means of both stochastic and deterministic models. In her contribution, Pasquali [46] focuses on stage-structured demographic models, in which growth of an individual is described by its physiological age, which is governed by a stochastic differential equation.

1.4 INTEGRATING POPULATION DYNAMICS AND POPULATION STRUCTURE

Matrix population models that integrate population dynamics and population structure are a power tool for investigating population dynamics [47]. It has also been long known that age specific effects have a profound influence on overall population dynamics [48]. Matrix projection model based on age specific characteristics of individuals [49;50;34], have become a means of characterizing populations and predicting their future behaviour [51]. In some areas of resource biology, the sampling programmes are mobilized in order to obtain data to build life tables from which population dynamics models may be developed. The primary objectives of population dynamics modelling are twofold; to give insight into the biological mechanisms operating in the system being modelled and to produce a model of community interactions which predicts changes in abundances (numbers of community species). In recent years, there has been a remarkable expansion in the application of matrix models. Quantitative demographic analysis should be used more frequently to advise management, according to population ecologists [52;53;54]. The Discrete Leslie Matrix Model [55;34;56] has been used to analyze the growth in age structured populations. Matrix population models that integrate population dynamics and population structure are a power tool for investigating population dynamics [47]. It has also been long known that age specific effects have a profound influence on overall population dynamics [57].

In this study, the Discrete Leslie Matrix Model [55;34;56] has been used to analyse the growth in age structured population of *Euphorbia hirta* Linn, an invasive species in agrosystems with the following objectives:

i) to study the temporal dynamics of target weed population

- ii) to investigate the transient dynamics and asymptotic characteristics of the population
- iii) to project the structure and size of future populations and
- (iv) to suggest the critical stages in the life cycle of the weed for management programme

2 Material and Methods

2.1 The Species

Euphorbia hirta Linn. selected for the present study, commonly known as Pill-bearing spurge, is a member of the Euphorbiaceae family. It is a little annual herb that is propagated by seeds. The species is currently found throughout tropical and subtropical regions, having originated in Tropical America. It is commonly observed inhabiting paths, roadside vegetation, grasslands, banks of watercourses, and open waste areas [58]. Due to increased trade, tourism, industry growth, transportation, technology advancements, and rising rates of urbanization, the invasion of species has expanded significantly [59]. According to Pauchard et al. [60], the invasion of various regions by alien species increase pressure on the natural environment.

Euphorbia hirta is an important agrestal weed with special affinity for paddy tracts, irrigated and garden crops. It has been documented that *Euphorbia hirta* infestations occur in rice, mung-sesame system, chilli, maize, and mustard (*B. juncea*) crops [61;62; 63]. Aqueous extract of *E. hirta* at high concentration was found to impair the growth of maize and wheat seedlings, delayed germination, reduced chlorophyll and wheat protein content [63]. It exerts allelopathic effect on crops like potato, sugarcane, maize and sorghum by competing with native plants and discouraging grazing near it, the plant reduces forage production and interferes with pasture/rangeland and livestock [64]. This directly affects the land's suitability for livestock grazing.

Weeds on rangelands have an adverse effect on the livestock industry by reducing forage supply and its quality, obstructing grazing, poisoning animals, raising the expense of managing and producing livestock, and decreasing land value [65;66]. In pastures and rangelands, *E. hirta* easily outcompetes desirable vegetation [67]. After establishing itself in pasture and rangeland habitats, it tends to displace all other vegetation [68;69]. Found widely in moist and dry environments, *E. hirta* creates through allelopathy, essentially a single species stand by releasing the

flavonoid compound Kaempferol glucuronide, the plant is poisonous to animals and presents a significant risk to the productivity of livestock on open rangelands [70]. *E. hirta* is selected for the present study.

2.2 Study Area:

The study was conducted around Imphal (24°75' N latitude and 93°85' E Longitude at an elevation of 782 m MSL) at Imphal valley, Manipur, North East India. The average maximum temperature ranges from 25.1°C-31.1°C (May - June) while the minimum temperature ranges from 11.8°C to 19.4°C. (December-January), the average annual rainfall being 1470 mm per annum. Humidity is highly variable during different seasons ranging from 45 to 100%.

2.3 Demographic Analysis:

Over two-year period, field study was conducted on pure natural population of forty randomly chosen 0.25m² plots demarcated for studying the demographic parameters. The number of individuals in 4 different functional stages called age groups were recognized in the field conditions and marked. These 4-age group comprised of seedling, juvenile, flowering adults and fruiting adults. Record of the individual plants in the above said area for each stage was made. New individuals becoming established at the time of each census were identified as per age group. The number of individuals in each age group were censused at monthly intervals and the observations were continued for more than 2 years (December 2020-December 2022). To assess the impact of nutrient supply on seedling emergence, soil samples were taken at 2-6 cm depth from ten adjacent plots and analyzed for Soil Moisture, pH, Organic Carbon, Total Nitrogen, Available Phosphorus and Exchangeable Potassium as per the standard methods given in Misra [71].

2.4 The Model

The equation,

$$Av = \lambda v$$

indicates that A = square matrix

v = column vector and

λ = scalar.

For every Eigen value λ, there is an associated Eigen vector v. The dominant Eigen value gives the rate at which the population size increases.

The Leslie matrix model take the general form of $N(t+1) = A.N(t)$.

Where, $N(t+1)$ and $N(t)$ are vectors representing age or stage class distribution of individuals at time t+1 and t, A is the matrix defining the age or stage specific survival and fecundity values. The rate of emigration and immigration of propagules were assumed to be equal. The model was thus simplified. The asymptotic population growth rate λ is given by the dominant Eigen values of A₁ the stable age/stage distribution by corresponding Eigen vector. Rate of decay or death rate was designated as -μ and calculated at each time interval.

2.5 General mathematical formulation of the population dynamics Model:

As the number of individuals changes with time t,

$$\frac{dN}{dt} \propto N$$

$$\frac{dN}{dt} = m.N$$

$$\frac{dN}{N} = m.dt$$

$$\int \frac{dN}{N} = m. \int dt$$

$$\text{Log } N = m.t + C$$

$$\text{at } t = 0, N = N_0$$

$$\text{Since } \log N_0 = C$$

$$\text{Log } N = mt + \log N_0$$

$$\text{Log } \frac{N}{N_0} = mt$$

So,

$$N = N_0 e^{mt}$$

$$\text{Put } e^{mt} = \lambda = \text{growth rate}$$

$$N = \lambda N_0$$

$$\text{or } N(t+1) = \lambda N(t)$$

where t = the initial observation year

t +1 = a time period of one year after the initial observations were recorded

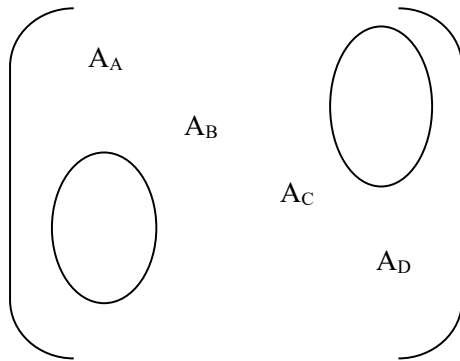
When λ equals 1, the population size is constant but increases or decreases when λ is more than 1 or less than 1 respectively.

2.6 Application of the Model

The age distribution vector were derived from log graphs where the abscissae demarcated the length of span for each age group scaled in such a way that the length of life span for seedling stage equals one unit and on the ordinates the density of different age groups were plotted.

The contribution to the population at time $t+1$ can be obtained by multiplying N_t by the following matrix comprised of $a \times a$ as submatrices,

where, A_A to A_D represent the survivorship of 4 stages during the time interval.



2.7 Groups and Cohorts

In weed population 4 stage classes were recognized which have specific longevity. Taking seedling state life span as the unit, all the other 3 stage classes were subdivided into a number of cohorts for projecting population structure of future time as it was assumed that the individuals in a particular ontogenetic stage with the experimentally derived longevity can be splitted over cohorts. Thus, taking the unit time interval of 5 days in case of *Euphorbia hirta* the total life span of the weed can be splitted into 17 cohorts. Thus, the column vector $N_{(t)}$ comprised of age classes (derived graphically on log scale).

2.8 Survival Coefficients

Survival coefficients were calculated for all stage classes and are regarded as the proportion of individuals borne at a given time, actually the survival schedule of the individual takes into consideration the number of individuals surviving to a particular stage. To obtain survival coefficients, it was seen that how many individuals borne at a particular time survived the first interval of time, how many the second, how many the third and so on until no more were alive. Thus probabilities of surviving from one age group to the next were calculated. The population structure at $t+1$ time can be derived by multiplying $N(t)$ with a block diagonal matrix with diagonal submatrices A_A , A_B , A_C and A_D where A_A to A_D indicated the survivorship of four stage classes during the time interval through 17 age groups in *Euphorbia hirta*. The number of groups in seedling, juvenile, young and mature stage classes were 1,2,5 and 9 for *E. hirta*.

Based on age transitions the matrices were computed and equal to the proportion of plants that

were in the j^{th} age class at time t that entered the i^{th} age class at time $t+1$.

2.9 Possibility of Seed Bank

The area had pure standing populations of the weed studied. The rate of migration and entrance (immigration) of propagules were assumed to be constant taking into consideration the fact that propagules had equal chance of dispersal in the field. The model was thus simplified as effect of wind velocity etc. on dispersal was not seen. The effect of the dispersal agencies if any, was supposed to be counterbalanced as the studied areas were located within the pure stands of the weed population.

2.10 Projecting Population Structure

Leslie Matrix [56] modified after Gupta [71] describing the contribution of each age group to every other group during the time interval ($t, t+1$) was employed with a column vector $N(t)$ including the number of individuals in each group. Mostly emergent seedling population densities for 2021 were utilized as initial values in conventional Leslie Matrix form for simulation, the projected population structure and number of individuals in different age group was compared with the observed data in the field for the whole year. Relationship between seedling emergence and edaphic variables was regressed.

3 RESULTS AND DISCUSSION

3.1 Patterns of Growth

With experimental results available on *E. hirta*, it was found that the time of onset of flowering was 16 days which continued for 25 days, the maximum fruiting period was 45 days, the limit age of individuals involved in reproduction was 41 days and the life expectancy was 11 weeks.

3.2 Fluctuations in Population

Dynamics of *E. hirta* population for 2021 and 2022 are reflected in Fig 1. In *E. hirta*, the fluctuation range for seedling population varied from 47.3% to 37.5% in 2021 whereas 54.36% to 41.97% in 2022, for juvenile population, the range was 36.25% - 26.31% in 2021 and 36.56% to 26.1% in 2022, for flowering adult population, the range was 23.45% to 16.79% in 2021 and 24.13% - 15.15% in 2022, for flowering-fruiting adult, population range was from 7.89% - 1.92% in 2021 to 3.70% - 1.94% in 2022 respectively (Fig1).

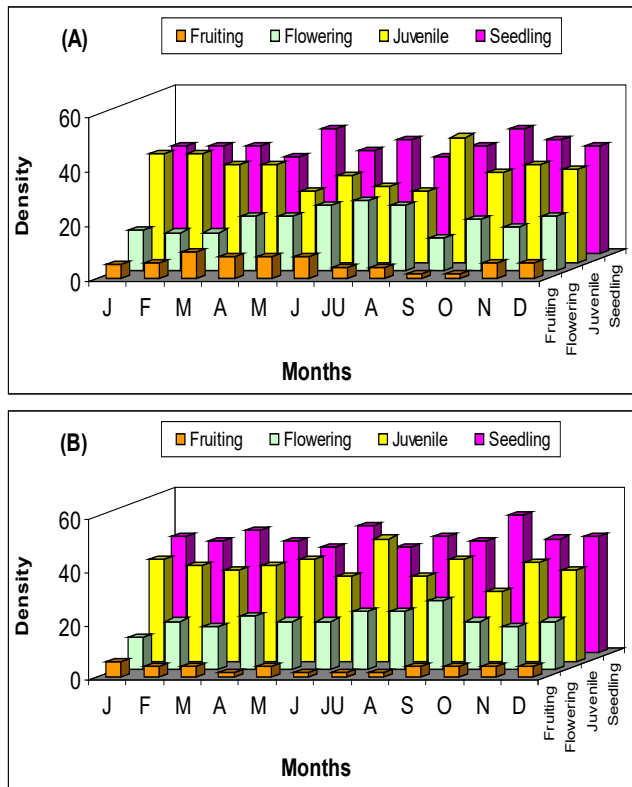


Fig.1: Dynamics of *Euphorbia hirta* population for (A) 2021 and (B) 2022((Density-Individuals 0.25 m⁻²).

3.3 Recruitment, Nutrient status and seedling emergent population

The outbreak in seedling population was obtained in December-January months in *E. hirta*.

An invasive species establishes when repeated reproduction and survival of individuals result in a population capable of maintaining itself in the wild [72]. Survival and reproduction rely on many abiotic (nonliving) factors that can either promote or hinder invasive species establishment.

The multiple linear regression derived was $Y = -221.07 + 1.69 X_1 - 16.51 X_2 + 189.5 X_3 + 820.73 X_4 - 5.67 X_5 + 0.59 X_6$

Where Y = seedling establishment

Numericals - 221.07 is the constant whereas other numerals are Regression coefficients.

X1- Soil Moisture, X2-pH, X3-Soil Organic Carbon, X4- Total Nitrogen, X5- Available Phosphorus and X6- Exchangeable Potassium

The multiple linear regression in *E. hirta* showed that the species responses were linear to the abiotic factors studied on seedling establishment and were significant both at 0.05 and 0.01 levels. ($r^2=0.970$, $F=26.948$ at df 6, 5).

The relative maximum percentage contribution on seedling establishment was exhibited by Exchangeable Potassium (59%) followed by pH (39.3%) in *E. hirta*.

3.4 Age Structure and Age Pyramids

The age structure into 4 age groups viz. seedling, juvenile, flowering adult and flowering –fruiting adult was expressed as a combination of total annual density in the studied area of corresponding age group resulting in age pyramid. The fluctuations of population age structure in the two years revealed the fate of various cohorts in *E. hirta* (Fig 2).

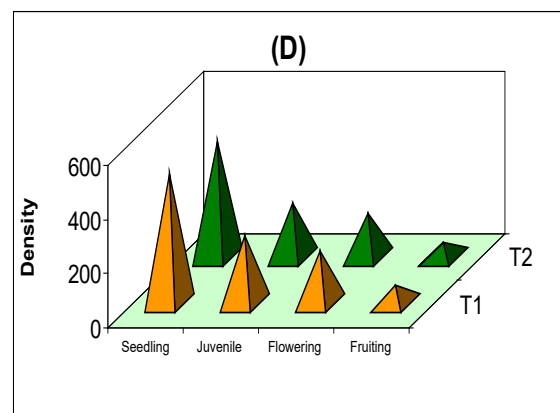


Fig 2 Age pyramid for populations of *Euphorbia hirta*

3.5 Age Specific Survival and Mortality

The monthly survival and mortality percentages for seedling and juvenile age group of weed population for a period of 2 years were computed. During the study period, the maximum survival percentage by seedling population was exhibited in the month of June (96%) and July (95.56%) in *Euphorbia hirta* for the year 2021 and 2022 respectively. The maximum mortality % in seedling stage was exhibited in the months of April (55.56%) and September (46.67%) in the species for the year 2021 and 2022.

Whereas the population in juvenile stage showed the maximum survival percentage in the months of June (86.36% and 77.77%) for the year 2021 and 2022 respectively. The maximum mortality percentage in the months of May (64.04% and 60.53%) for 2021 and 2022 respectively.

3.6 Survival Coefficients

The survival coefficients showed a flux in various age groups within the same species. Fig 3 reveals the survival coefficients for different months on the basis of average of two years value in 4 age groups of weed population in Manipur. They are the possible transitional probabilities and reflected the temporal variation in population structure.

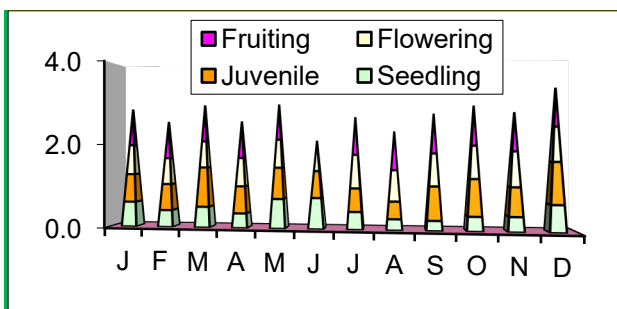


Fig 3 Survival coefficients for *Euphorbia hirta* L. for four different age groups

The lower value of survival coefficients was obtained for seedling population in *Euphorbia hirta* in September that revealed the risky nature of this stage in the life span.

3.7 Growth Rate (λ)

The weed population of *Euphorbia hirta* exhibited an annual percentage gain of 10.55%. The number of new individuals in 2 years (2021 and 2022) revealed the fate of various cohorts and subcohorts in the weed species. The average survival and mortality percentages were 75.65% and 24.35% respectively *hirta* exhibited positive density dependent correlation, growth rate exceeded value 1.0 in the species, the annual decay rate was 0.235 (Fig 4).

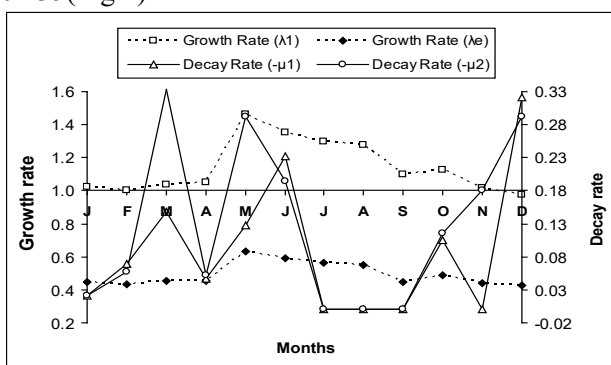


Fig 4 Growth Rate (λ) and Decay Rate ($-\mu$) for the populations of *Euphorbia hirta*

3.8 Population behavior

Gupta (71;73;74;75;76) made attempts to combine age and structure stage into one and analysed the

demography of herbaceous annuals by matrix model.

The Leslie matrix model take the general form of $N(t+1) = A.N(t)$,

where, $N(t+1)$ and $N(t)$ are vectors representing age or stage class distribution of individuals at time $t+1$ and t , A is the matrix defining the age or stage specific survival and fecundity values. The rate of emigration and immigration of propagules were assumed to be equal. The model was thus simplified. The asymptotic population growth rate λ is given by the dominant Eigen values of A_1 the stable age/stage distribution by corresponding Eigen vector. Rate of decay or death rate was designated as $(-\mu)$ and calculated at each time interval.

The age distribution vector was derived from log graphs where the abscissae demarcated the length of span for each age group scaled in such a way that the length of life span for seedling stage equals one unit and on the ordinates the density of different age groups were plotted in all the herbaceous annuals populations. It was assumed that individuals in a particular ontogeny with experimentally derived longevity can be splitted into subcohorts

The Column Vector $N(t)$ comprised of: $N(t) = (68.0, 56.0, 43.0, 35.0, 27.2, 21.0, 14.1, 8.0, 7.0, 5.3, 5.0, 4.0, 3.8, 3.6, 3.2, 3.0, 2.8)$.

In *E. hirta* through 17 cohorts (with Seedlings life span 5 days)

The contribution to the population at time $t+1$ can be obtained by multiplying N_t by the matrix defining the age or stage specific survival matrix. For simulations the monthly emergent seedling population for the initial year were analysed in case of all plants and multiplied with survivorship coefficient.

3.9 Population Projection

The age structure of projected population for 2022 revealed its resemblance with observed age structure for 2022 population in field conditions. It indicated that the projection matrix is satisfactory in differentiating three age specific characteristics of cohorts. In both the cases of projected and observed populations, the largest category of individuals was composed of seedlings followed by juveniles, then flowering stage and lastly fruiting adult stage plants. Thus, a remarkable similarity between projected and observed age group distribution curves were noticed (Fig.5).

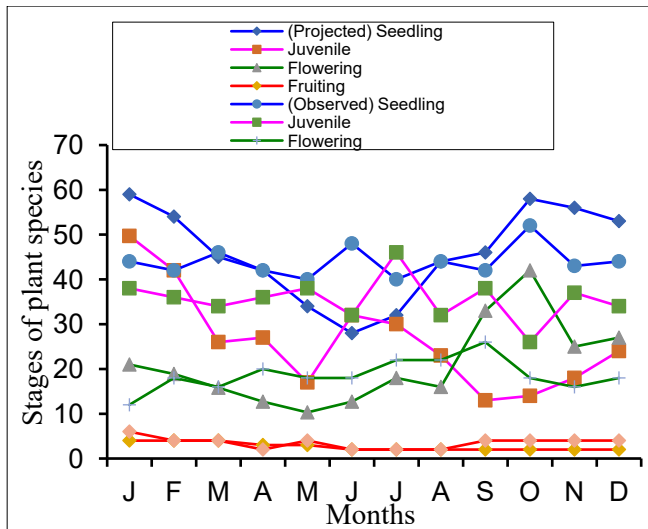


Fig 5 Projection for *Euphorbia hirta* L Population (Leslie Matrix Model)

The simulation indicated a striking similarity between observed and projected populations. It indicated that the projection matrix was satisfactory in differentiating the age specific characteristics of cohorts and sub-cohorts. The present study also substantiated the earlier works [71;73;74;75;76].

The matrix model provided a path elucidating the dynamics of population in which reproduction and survival coefficient for seedling stage suggested it a critical stage in life cycle of *E. hirta*. Thus, the lower value of survival coefficients obtained for seedling population in the month of September revealed that the mortality risk was highest at seedling stage suggesting to have the control measures at seedling stage. In, *Portulaca oleracea* [75;76], the seedling stage was also regarded as risky due to minimum survival coefficient value. Earlier works [71;73] on *Parthenium hysterophorus* revealed flowering stage a critical stage in the life cycle of this notorious weed that approved that the effective control of the weed may be brought about by regulating it at flowering stage.

Similarly, for *Tridax procumbens*, Linn., flowering was obtained as a sensitive stage. Thus, it appears that a division of resources between vegetative and reproductive organ is important for it [74;75], whereas in *Bidens pilosa* survival co-efficient of juvenile stage was found to be low indicating that competition for nutrient and space (Dense turf) is an important factor in regulating it [74;75].

The study reveals that seedling stage is one of the riskiest phases in the life-history of *E. hirta*. It is considered to be the most vulnerable stage in the life

of the plant [77]. Fenner & Thompson [78] owe this to reduction in biomass, if even reduced in small amount may lead to the death of the plant. Seedlings face threats to their establishment from natural enemies to resource limitations and insufficiencies in sites suitability [79]. Seedling recruitment can depend on abiotic and/or biotic variation at very small scales (e.g. meters or weeks), yet be a major demographic driver of community dynamics and species distributions [80;81] Burkey and Stenseth [82] demonstrated with a seasonal model how the value of the resource to each individual may be reduced due to its patchy distribution. It was concluded that the difference among population of various weed species depend on the difference of their transition probabilities of matrices in the same year and on the difference in stochastic processes. As also observed by Aberg [83], the difference was reflected in value of λ and $-\mu$ rates.

4 Conclusion

Matrix population models offer a tool for identifying the demographic processes that have the greatest impact on population growth rates in order to better understand population dynamics and potential management strategies for invasive plant species. I demonstrate that the population under study has a higher growth rate and is demographically stable through the examination of asymptotic growth rate and demographic parameters. Population declines of invasive species may be achieved by focusing control on demographic processes (survival) to target for reductions in population. Reducing the survival and growth of the species would have the biggest impact on lowering the overall population growth rate. Invasive species can experience population losses by concentrating management on demographic processes. Applying density thresholds to a transient invasive species (*E.hirta*), I examined how population density affected population projection and offered for management recommendations.

As invasive species management is concerned with maximizing mortality and lowering the reproduction and minimizing the loss of resources resulting from invasive species competition with crops, it is suggested to target the population for management measures at seedling stage so as to eliminate the probability of reaching the adult stage thereby setting seeds for further infestation and spread. Our results provide a first evaluation of the demography of *E. hirta*, a species of economic concern, and

provide the first structured population model of a representative of the *Euphorbiaceae* family, thus contributing to our global knowledge on plant population dynamics.

5 Recommendations and Future work

Invasive species, a global threat due to climate change and human disturbances, require urgent management strategies. Early detection and swift action are crucial for eradicating these species which disrupt ecosystems, impact adversely biodiversity, introduce diseases, and cause financial burdens. Matrix models provide the tools that can identify high-risk stage in the life cycle, prioritizing management and population eradication.

Recruitment from seed banks impacts population structure in time and space, with age structure resulting from long-term persistence in the seed bank. Changes in age structure and population size significantly influence demography, necessitating inclusion of age-structured seed bank dynamics in demographic models that I intend to include in my future work.

Acknowledgement

The author wholeheartedly thanks the three anonymous reviewers whose criticisms and constructive suggestions helped to a great extent in improvement of the present paper. Author also acknowledges the Head, Centre of Advanced Study, Department of Life Sciences, Manipur University for the facilities.

References:

- [1]. CBD, *Alien species that threaten ecosystems, habitats or species*. 2002, pp. 249-261.
<https://www.cbd.int/doc/decisions/cop-06-dec23-en.pdf>
- [2]. Gaertner, M, Breeyen, A.D, Hui, C, Richardson, D.M., Impacts of alien plant invasions on species richness in Mediterranean-type ecosystems: a meta-analysis. *Progress in Physical Geography*, Vol 33, 2009, pp. 319–338.
- [3]. Sardans, J, Bartrons, M, Margalef, O, Gargallo-Garriga, A, Janssens, I.A, Ciais, P., Plant invasion is associated with higher plant-soil nutrient concentrations in nutrient-poor environments. *Global Change Biology*, Vol 23, 2017, pp. 1282-1291.
- [4]. Funk, J.L., The physiology of invasive plants in low-resource environments. *Conservation Physiology*, Vol 1, 2013, pp. 1-17. DOI: 10.1093/conphys/cot026.
- [5]. Kohli, R.K, Jose, S, Singh, H.P, Batish, D.R., *Invasive Plants and Forest Ecosystems*. New York, CRC Press, Taylor and Francis Group, 2009.
- [6]. Early, R, Bethany, A.B, Dukes, J, Joshua, J.L, Julian, D.O, Dana, M.B, et al Global threats from invasive alien species in the twenty-first century and national response capacities. *Nature Communications*, Vol 7, 2016 pp. 12485. DOI: 10.1038/ncomms12485.
- [7]. Clout, M.K, DePoorter, M.D., International initiatives against invasive alien species. *Weed Technology*, Vol 19, 2005, pp. 523-527.
- [8]. D'Antonio, C., Meyerson, L. A., Denslow, J., Research Priorities for the Next Decade. *Conservation Biology* eds. Soulé, M. E. & Orians, G. H. Island, Washington, DC, 2001, pp. 59- 80.
- [9]. Pejchar, L, Mooney, H.A., Invasive species, ecosystem services and human well-being. *Trends in Ecology and Evolution*, Vol 24, No.9, 2009, pp. 497- 504.
- [10]. Ricciardi, A, Neves, R.J, Rasmussen, J.B., Impending extinction of North American fresh water mussels (Unionoida) following the zebra mussel (*Dreissena polymorpha*) invasion. *Journal of Animal Ecology*, Vol 67, 1998, pp. 613-619.
- [11]. Vitousek, P.M., Biological invasions and ecosystem processes: towards an integration of population biology and ecosystem studies. *Oikos*, Vol 57, No1, 1990, pp. 7–13.
- [12]. Mazar, T., Doropoulos, C., Schwarzmuller, F., Gladish, D. W., Kumaran, N., Merkel, K., Global mismatch of policy and research on drivers of biodiversity loss. *Nat. Ecol. Evol.* Vol 2, 2018, pp 1071–1074. doi: 10.1038/s41559-018-0563-x
- [13]. Shiferaw H, Bewket W, Alamirew T, Zeleke G, Teketay D, Bekele K, Schaffner U, Eckert S., Implications of land use/land cover dynamics and *Prosopis* invasion on ecosystem service values in Afar Region, Ethiopia. *Sci Total Environ* Vol.675, 2019, pp.354–366.
- [14]. Bansal, S, Lishawa, S.C, Newman, S, et al., *Typha* (cattail) invasion in north American wetlands: Biology, regional problems, impacts, ecosystem services, and management. *Wetlands*, Vol 39, 2019, pp.

- 645-684. DOI: 10.1007/s13157-019-01174-7.
- [15]. Fleming, J.P, Wersal, R.M, Madsen, J.D, Dibble, E.D., Weak non-linear influences of biotic and abiotic factors on invasive macrophyte occurrence. *Aquatic Invasions*, Vol 16, No2, 2021, pp. 349-364.
- [16]. Li, W, Luo, J, Tian, X, et al., A new strategy for controlling invasive weeds: Selecting valuable native plants to defeat them. *Scientific Reports*, Vol 5, 2015, pp.11004. DOI: 10.1038/srep1100.
- [17]. Kriticos, D.J, Sutherst, R.W, Brown, J.R, Adkins, S.W, Maywald, G.F., Climate change and the potential distribution of an invasive alien plant: *Acacia nilotica* subspecies *indica* in Australia. *Journal of Applied Ecology*, Vol 40, No1, 2003, pp. 111-124.
- [18]. Hastings, A., Cuddington, K., Davies, K. F., Dugaw, C. J., Elmendorf, S., Freestone, A., Thomson, D., The spatial spread of invasions: New developments in theory and evidence. *Ecology Letters*, Vol 8, No1, 2005, pp 91–101. <https://doi.org/10.1111/j.1461-0248.2004.00687>.
- [19]. Büyüktaktakın, I. E, Haight, R. G, A review of operations research models in invasive species management: State of the art, challenges, and future directions. *Annals of Operations Research*, Vol 271, No.2 ,2018, pp. 357–403. <https://doi.org/10.1007/s10479-017-2670-5>
- [20]. Baker, C. M., Bode, M., Recent advances of quantitative modeling to support invasive species eradication on islands. *Conservation Science and Practice*, Vol3, No2,2021.<https://doi.org/10.1111/csp2.246>
- [21]. Liu, C., Wolter, C., Xian, W., & Jeschke, J. M. (2020). Most invasive species largely conserve their climatic niche. *Proceedings of the National Academy of Sciences USA*, 117, 23643–23651 <https://doi.org/10.1073/pnas.2004289117>
- [22]. Bogdan A. , Levin S C , Salguero-Gómez R, Tiffany M. Knight, Demographic analysis of an Israeli *Carpobrotus* population. *Plos One*, Vol16, No.4,2021,e0250879. doi: 10.1371/journal.pone.0250879.
- [23]. Chung H I, Choi Y, Yoo Y, Engler R, Lee K and S W Jeon Integrated spatial model-based evaluation methodology for optimal invasive species management: common ragweed in the Republic of Korea *Environ. Res. Lett.* Vol.17 ,2022, pp 034-047 <https://doi.org/10.1088/1748-9326/ac4dc7>
- [24]. Guetling C H · Jones L.C. Strand E K. and D W. Morishita, Piaskowski J and Timothy S. P. Two invasive Hieracium species' potential distributions within the Greater Yellowstone Ecosystem were defined using invasion susceptibility models and habitat typing · *Biol Invasions*, Vol 25, 2023, pp 2231–2248 <https://doi.org/10.1007/s10530-023-03037-z>
- [25]. Rew LJ, Maxwell BD Stratified random sampling method. In: Rew LJ, Pokorny ML (eds) Inventory and survey methods for nonindigenous plant species. Bozeman, MT Montana State University Extension, 2006 pp 49–55of *Applied Ecology*, Vol55, pp526-538. <https://doi.org/10.1111/1365-2664.13025>
- [26]. Davis AJS, Groom Q, Adriaens T, Vanderhoeven S, De Troch R, Oldoni D, Desmet P, Reyserhove L, Lens L and Strubbe D, Reproducible WiSDM: a workflow for reproducible invasive alien species risk maps under climate change scenarios using standardized open data. *Front. Ecol. Evol.* Vol 12,2024, pp.1148895. doi: 10.3389/fevo.2024.1148895
- [27]. Liu, Y, Oduor, A.M.O, Zhang, Z, Manea, A, Tooth, I.M., Leishman, M.R, et al., Do invasive alien plants benefit more from global environmental change than native plants? *Glob Chang Biol* Vol 23, No8, 2017 pp. 3363–70. 10.1111/gcb.13579
- [28]. Caswell, H., *Matrix Population Models: construction, analysis, and interpretation*. 2001 2nd ed. Sunderland, Massachusetts: Sinauer Associates; [Google Scholar]
- [29]. Easterling, Michael R., Stephen, P. Ellner, P.M.D., Size-specific sensitivity: applying a new structured population model. *Ecology*, Vol 81, No3, 2000, pp. 694–708. [Google Scholar]
- [30]. Ramula, S, Knight, T.M., Burns J.H, Buckley, Y.M., General guidelines for invasive plant management based on comparative demography of invasive and native plant populations. *J Appl Ecol*, Vol 45, No 4, 2008, pp. 1124–33. [Google Scholar]
- [31]. Silvertown J, Franco M, Menges E. Interpretation of elasticity matrices as an

- aid to the management of plant populations for conservation. *Conserv Biol.*, Vol 10, No 2, 1996, pp 591–597. [Google Scholar].
- [32]. Sagar, G. R., Factors controlling the size of plant populations. *Proceedings 10th Brit. Weed Control Conf*, 1970, pp. 965-79.
- [33]. Sagar, G.R., Mortimer, A.M., An approach to the study of the population dynamics of plants with special reference to weeds. *Applied Biolog.*, Vol 1, 1976, pp. 1-43.
- [34]. Leslie P.H., On the use of matrices in certain population mathematics. *Biometrika*, Vol 33, 1945 pp. 183–212. doi: 10.2307/2332297.
- [35]. Lefkovich, L.P., Study of population growth in organisms grouped by stages. *Biometrics*, Vol 21, 1965, pp. 1–18. doi: 10.2307/2528348.
- [36]. Caswell, H., General formula for sensitivity of population-growth rate to changes in life-history parameters. *Theoretical Population Biology*, 1978, pp 215–230. DOI 10.1016/0040-5809(78)90025-4.
- [37]. De Kroon, H, Van Groenendael, J, Ehrlen, J., Elasticities: a review of methods and model limitations. *Ecology*, Vol 81, 2000, pp. 607–618.
- [38]. Cochran, M.E, Ellner, S., Simple methods for calculating age-based life-history parameters for stage-structured populations. *Ecological Monographs*, Vol 62, 2000, pp. 345–364. DOI 10.2307/2937115
- [39]. Brault, S, Caswell, H., Pod-specific demography of killer whales (*Orcinus orca*). *Ecology*, Vol 74, 1993, pp. 1444–1454. DOI 10.2307/1940073.
- [40]. Caswell, H, Brault, S, Read, A.J, Smith, T.D., Harbor porpoise and fisheries: an uncertainty analysis of incidental mortality. *Ecological Applications*, Vol 8, 1998, pp. 1226–1238. DOI 10.1890/1051-0761(1998)008[1226:HPFAUJ]2.0.CO;2.
- [41]. Crouse, D.T., Crowder, L.B, Caswell, H., A stage-based population-model for Loggerhead sea-turtles and implications for conservation. *Ecology*, Vol 68, 1987, pp. 1412–1423. DOI 10.2307/1939225.
- [42]. Crowder, L.B, Crouse, D.T, Heppell, S.S, Martin, T.H. 1994. Predicting the impact of turtle excluder devices on Loggerhead sea-turtle populations. *Ecological Applications*, Vol 4, 1994, pp. 437–445. DOI 10.2307/1941948
- [43]. Fujiwara M, Caswell H., Estimating population projection matrices from multi-stage mark-recapture data. *Ecology*, Vol 83, 2002, pp. 3257–3265. DOI 10.2307/3072076.
- [44]. Salguero-Gomez, R, Plotkin, J.B., Matrix dimensions bias demographic inferences: implications for comparative plant demography. *American Naturalist*, Vol. 176, 2010, pp. 710–722. DOI 10.1086/657044.
- [45]. Lebreton, J.D., Age, stages, and the role of generation time in matrix models. *Ecological Modelling*, Vol 188, 2005, pp. 22–29. DOI 10.1016/j.ecolmodel.2005.05.003.
- [46]. Pasquali S.A., Stage structured demographic model with “no-regression” growth: The case of constant development rate. *Physica A: Statistical Mechanics and its Applications*, Vol 581, 2021, pp 126200.
- [47]. Caswell, H., *Matrix Population Model-1*. Sinauer Asso. Inc. Publ. 1989. Sunderland, Massachusetts.
- [48]. Lotka, A.S, *Elements of Physical Biology*, 1956. Williams and Wilkins, Baltimore, Dover Publications, New York.
- [49]. Lewis, D., The evolution of sex in flowering plants. *Biological Reviews* Cambridge Philosophical Society, Vol 17, No1, 1942, pp. 46-67.
- [50]. Lewis, E. R., Delay-Line Models of Population Growth. *Ecology*, Vol 53, No 5, 1972, pp. 797-807.
- [51]. Keyfitz, N, *Population of The World and Its Regions 1975-2050*. (WP-77-7 2361 International Institute for Applied Systems Analysis, 2361 /Luxembarg, Austria, 2022).
- [52]. Schemse, D.W, Husband, B.C., Ruckelshaus, M.H., Goodwillie, C, Parker, I.M, Bishop, J.B, Evaluating approaches to the conservation of rare and endangered plants. *Ecology*, Vol 7, 1994, pp 584-606.
- [53]. Bakker, V. J, Doak, D.F, Population viability management: ecological standards to guide adaptive management for rare species. *Front Ecol Environ*, Vol 7, 2009, pp 158-165.
- [54]. Ramula, S, Buckley, Y., Demographic models as a tool to understand and manage invasive plant species Conference Paper: 94th *ESA Annual Convention* 2009.
- [55]. Lewis, D., The evolution of sex in flowering plants. *Biological Reviews*

- Cambridge Philosophical Society*, Vol 17, No.1, 1942, pp. 46-67.
- [56]. Leslie, P.H., Some further notes in the use of matrices in population mathematics. *Biometrika*, Vol 35, 1948, pp. 213-245.
- [57]. Lotka, A.S., *Elements of Physical Biology*, 1956. Williams and Wilkins, Baltimore. Dover Publications, New York.
- [58]. Sandeep, B.P, Nilofar, S.N, Chandrakant, S.M., Review on phytochemistry and pharmacological aspects of *Euphorbia hirta* Linn. *J Pharma Res Health Care*, Vol 1, 2009, pp. 113-133.
- [59]. Ullah, R., Khan, N., Rahman, A., Comparison of invaded and non-invaded sites: a case study of rough cocklebur (*Xanthium strumarium* L.) An alien invasive species. *Appl. Ecol. Environ. Res*, Vol 19, No.2, 2022, pp. 1533–1548. doi: 10.15666/aecer/1902_15331548.
- [60]. Pauchard, A., Kueffer, C., Dietz, H., Daehler, C. C., Alexander, J., Edwards, P. J., Seipel, T., et al., Ain't no mountain high enough: plant invasions reaching new elevations. *Front. Ecol. Environ*, Vol 7 No 9, 2009, pp. 479–486. doi: 10.1890/080072
- [61]. Sandhu, K. S., Singh, T., Singh, S., Weed composition of maize (*Zea mays*) fields in Punjab. *Indian J. Weed Sci*, Vol 31, No 1/2, 1999, pp. 18-24.
- [62]. Kumar, S., Sindhu, G., Tejasvi, A., Screening of chilli varieties for management of summer weeds. *Allelopathy J.*, Vol 24, No 1, 2009, pp. 143-156.
- [63]. Jabeen, N., Ahmad, M., Possible allelopathic effects of three different weeds on germination and growth of maize (*Zea mays* L.) cultivars. *Pak. J. Bot.*, Vol 41, No 4, 2009. pp.1677-1683.
- [64]. Huenneke, L. F., Ecological impacts of plant invasion in rangeland ecosystems. In: Alien Plant Invasions: Increasing Deterioration of Rangeland Ecosystem Health Symposium, Phoenix. *Proceedings Phoenix AZ*, 1995. pp. 10-14.
- [65]. Ditomaso, J M., Invasive weeds in rangelands: Species, impacts, and management. *Weed Science*, Vol. 48, No 2, 2000, pp. 255-265.
- [66]. Messersmith, C. G., The leaf spurge plant. *N.D. Farm Res*, Vol 40, No 5, 1983, pp. 3-7.
- [67]. Selleck, G. W., The antibiotic effect of plants in laboratory and field. *Weed Sci*, Vol 20, No 2, 1972, pp. 189-194.
- [68]. Steenhagen, D. A., Zimdahl, R. L., Allelopathy of leafy spurge (*Euphorbia esula*). *Weed Sci*, Vol 27, No 1, 1979, pp. 1-3,
- [69]. Selleck, G. W., Coupland, R. T., Frankton, C., Leaf spurge in Saskatchewan. *Ecol. Monogr*, Vol 32, No 1, 1962, pp. 1-29.
- [70]. Misra, R. *Ecology Workbook*. 1968. Oxford and IBM Publ. Co., New Delhi.
- [71]. Gupta A., Population Ecology of *Parthenium hysterophorus* Linn. in Herbaceous Communities of Central India (State – Maharashtra) *Evolutionary Morphology and Ecology*, Acad. of Sc. USSR, 1990, pp. 1-36.
- [72]. Blackburn, T. M., Pyšek, P., Bacher, S., Carlton, J. T., Duncan, R. P., Jarošík, V., Wilson, J. R.U., Richardson D.M. A proposed unified framework for biological invasions. *Trends Ecol Evol*. Vol 26, No.7, 2011 pp.333-339. doi: 10.1016/j.tree.2011.03.023.
- [73]. Gupta A., Population Dynamics studies in Congress Grass (*Parthenium hysterophorus* Linn.) *Proc. XV International. Botany Congress*, Yokohama, Japan; 1993
- [74]. Gupta A, Matrix Model population for two herbaceous species. *Int. J. Ecology and Environmental Sc.*, Vol 23, 1997, pp. 197-207.
- [75]. Gupta A., Matrix model in Ecology . In *Perspective in Ecology* 1999, pp. 197-203 Ed. A. Farina, Backhuys Publ. Leiden, NL.
- [76]. Gupta, A., A Matrix population model for *Portulaca oleracea* Linn. *Ecology, Environment and Conservation* , Vol 12, No 4, 2006 , pp. 637-642.
- [77]. Sttebbins, G.L., Relationships between adaptive radiation, speciation and major evolutionary trends. *Taxon*, Vol 20, No 1, 1971, pp. 3-16
- [78]. Fenner, M., Thompson, K. *The Ecology of Seeds* 2005, pp. 260. Cambridge University Press.
- [79]. Harper, J.L, Williams, J.T., and Sagar, G.R., The behaviour of seeds in soil: The heterogeneity of soil surfaces and its role in determining the establishment of plants

- from seed. *J of Ecology*, Vol 53, No 2, 1965, pp 273-286.
- [80]. Lebrija-Trejos, E., Reich, P. B., Hernandez, A., Wright, S. J., Species with greater seed mass are more tolerant of conspecific neighbours: A key driver of early survival and future abundances in a tropical forest. *Ecology Letters*, Vol 19, 2016, pp. 1071–1080.
- [81]. Stampfli, A., Zeiter, M., Mechanisms of structural change derived from patterns of seedling emergence and mortality in a semi-natural meadow. *Journal of Vegetation Science*, Vol, 19, 2008, pp. 563–574.
- [82]. Burkey, T.V., Stenseth, N.C., Population dynamics of territorial species in seasonal and patchy environments. *Oikos*, Vol 69, 1994, pp. 47-53.
- [83]. Aberg, P., A demographic study of two populations of the seaweed *Ascophyllum nodosum* *Ecology*, Vol 73, No 4, 1992, pp. 1473-1487.

Contribution of Individual Authors to the Creation of a Scientific Article (Ghostwriting Policy)

The author contributed in the present research, at all stages from the formulation of the problem to the final findings and solution.

Sources of Funding for Research Presented in a Scientific Article or Scientific Article Itself

No funding was received for conducting this study.

Conflict of Interest

The author has no conflict of interest to declare that is relevant to the content of this article.

Creative Commons Attribution License 4.0 (Attribution 4.0 International, CC BY 4.0)

This article is published under the terms of the Creative Commons Attribution License 4.0

https://creativecommons.org/licenses/by/4.0/deed.en_US